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Influence of reductant on the regeneration of SO₂-poisoned Pt/Ba/Al₂O₃ NO_x storage and reduction catalyst

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Abstract

A 10% Ba/Al₂O₃ in the presence and absence of 1% Pt was exposed to SO₂ at 673 K either in air or in nitrogen, and the ability to recover the NO_x storage capacity was compared after exposure to H₂, CO, or C₃H₆ at 673 K. Exposure to SO₂ led to adsorbed sulfite, surface aluminum sulfate, and bulk and surface barium sulfate formation, all of which reduced the NO_x storage capacity of the catalyst. The presence of Pt was required for bulk barium sulfate formation. A surface aluminum sulfate was formed from surface sulfite species to a limiting extent at 673 K but this was enhanced by heating in air to higher temperatures and was thermally stable up to 873 K. In the case of the Pt-containing catalyst, hydrogen at 673 K was able to completely reverse the detrimental effect of SO₂ exposure whereas H₂ exposure to SO₂-treated Pt-free catalyst was detrimental in terms of recovered NO_x storage capacity. The presence of oxygen in the feed stream during the rich regeneration period was significant as this allowed formation of water (and CO₂) which plays an important role in the regeneration process. © 2004 Elsevier Inc. All rights reserved.

Keywords: NSR catalysts; NOx trap; Pt/Ba/Al2O3; SO2 poisoning; Catalyst regeneration

1. Introduction

Storage and reduction catalysts (NSR) allow control of NO_x emissions from automobile sources while permitting operation under predominantly lean-burn conditions [1–8]. The concept is based on the storage of NO_x under lean conditions on an alkaline earth oxide component such as baria which is then released during intermittent rich/stoichiometric periods where the stored NO_x is released and reduced by H₂, CO, or HC over the noble metal component. The NO_x storage capacity of NSR catalysts are lowered as a function of time on stream as sulfate is accumulated on the storage component [3–8]. In cases where the extent of deactivation is relatively low, NO_x storage capacity may be readily regenerated under reducing conditions [9]. In more severe cases, even extended regeneration periods fail to recover the initial storage capacity [3,9]. This tran-

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sition in the ease by which sulfate may be removed may be associated with the transition from surface to bulk sulfates [4], although this is difficult to envisage as bulk barium carbonate still retains NO_x storage ability and it is generally accepted that NO_x storage is a surface rather than a bulk phenomenon [10,11]. It is also possible that sulfate particle size governs the ease by which this phase is decomposed [12]. It is generally considered that temperatures of between 723 and 823 K are required [5] and so failure to recover catalyst performance following desulfation may result from thermal degradation of the catalyst [5,6]. The difficulty in achieving complete regeneration following sulfate poisoning may also be linked to the regeneration procedure employed as these invariably involve exposure of the catalyst to a reducing atmosphere at high temperature [5] which may induce sulfide formation [3]. Although some of this sulfide is associated with Ba [3] there is also evidence that sulfide becomes associated with the metallic component [4,13] which therefore hinders further regeneration via reduction.

During the fuel-rich step, lean-burn engines emit CO (5-6%), H₂ (1-2%), and hydrocarbons (ppm levels), al-

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though the relative concentrations can be modified by engine management or insertion of a precatalyst. For example, hydrogen levels might be increased by water-gas shift or steam-reforming reactions. To shed further light on the deactivation and regeneration processes, the current study involved the use of CO, H₂, and C₃H₆ as reductants to examine the efficiency of trap regeneration for Ba/Al₂O₃ traps in the presence and absence of Pt which had been exposed to SO₂ either in air or in nitrogen. The objective was to determine whether sulphate buildup could be minimised, thereby limiting barium sulfate particle-size growth by selection of the appropriate conditions during the NO_x reduction cycles which would remove stored SO_x at each rich switch and thereby avoid progressive SO_x accumulation, loss of NO_x storage capacity, and the need to use high-temperature-rich excursions [5] to eliminate stable sulfates. The choice of NO2 rather than NO allowed an investigation of SO2 poisoning of support sites to be conducted by avoiding the need for Pt-catalysed oxidation as the latter might potentially become poisoned by SO_2 [4,13], leading to ambiguity regarding the origin of reduced NO_x uptakes.

2. Experimental

BaO/Al₂O₃ samples (10 wt% BaO) were prepared by precipitation of Ba(OH)₂ from a nitrate solution onto Degussa Aluminiumoxid C using an ammonia solution. The material was filtered, washed, dried at 363 K (16 h), and calcined in air at 773 K (2 h). A portion of the sample was wet impregnated with 1 wt% Pt (H2PtCl6) and excess water removed by heating while continually stirring. The resulting powder was dried overnight at 363 K and calcined in air $(100 \text{ cm}^3 \text{ min}^{-1})$ at 773 K (2 h). To distinguish between the influence of Pt and the influence of baria dispersion which is modified as a consequence of the addition of Pt precursor [14-16], a portion of the Ba/Al₂O₃ support was treated with HCl solution (pH 1.2) and then dried overnight at 363 K and washed with ammonia solution (pH 8.5). This sample is referred to as "HCl-treated support." It is clear [14] that the presence of this Cl may influence the adsorption properties of the support.

Samples for XRD purposes were calcined at 673 K and then exposed to 0.28 mmol of SO₂ at the same temperature followed by cooling in air to 298 K. One of the samples for XRD was regenerated with 3.35% H₂/4% air/balance N₂ mixture after SO₂ exposure at 673 K for 5 min and then cooling in air to 298 K. XRD patterns were obtained using a Philips PW1310 X-ray generator, using Cu-K α radiation ($\lambda = 1.542$ Å), coupled to a Philips PW1964 scintillation counter.

DRIFT spectra (4 cm⁻¹ resolution, 40 scans, MCT detector) were recorded using ca. 80-mg sample in a Harrick environmental cell. A gas blender fed the inlet gases, with exit gas analysis performed using on-line chemiluminescence (NO/NO₂) and Quadrupole mass spectrometer. Samples were calcined in situ (673 K, 1 h, 50 cm³ min⁻¹). SO₂

was injected as a series of $7 \times 1 \text{ cm}^3$ pulses either in flowing air or in N₂, and then the catalyst was either directly exposed to NO₂ or first exposed to a regeneration procedure which involved treatment in a flow containing H₂, CO, or propene under rich conditions. Gas mixtures contained 3.35% H₂ (or CO)/4% air/balance N₂ or 7.10% of 5% propene in N₂/4% air/balance N₂. Samples were exposed to NO₂ (750 ppm) at 673 K for 15 min and then cooled to 298 K in NO₂ before flushing in air for ca. 15 min to remove gaseous and weakly adsorbed forms of NO_x as indicated by the detector returning to background levels. A TPRS was performed between 298 and 873 K at 4 K min⁻¹ in 7.98% air/N₂ (50 cm³ min⁻¹) to determine the amount of stored NO_x.

3. Results

3.1. Sample characteristics

Ba/Al₂O₃ exhibited (2 θ) peaks at 24.1, 34.7, and 42.1° due to BaCO₃ and at 19.4, 22.0, and 28.0° due to BaAl₂O₄ [14,17] (Fig. 1a). There was no obvious change after exposure to SO₂ (Fig. 1b) (peaks at 2 θ = 44.6 and 38.3 are due to Al holder). Pt/Ba/Al₂O₃ displayed only features due to alumina, although the presence of minor, weaker diffraction lines due to BaCO₃ and BaAl₂O₄ cannot be ruled out. Following SO₂ treatment, additional peaks at 26.0, 28.7, and 42.5° appeared due to the presence of BaSO₄ [18]. These features were almost completely removed by H₂ treatment at 673 K (Fig. 1e).

3.2. Combined TPRS/FTIR experiments: Pt-containing catalysts

NO₂ was the dominant gas released by SO_x-free Pt/Ba/Al₂O₃ (Fig. 2a) giving maxima at 350, 530, and 729 K.

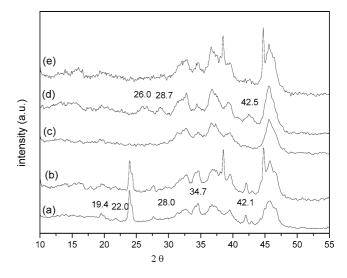


Fig. 1. XRD patterns of samples before and after SO_2 treatment at 673 K. Ba/Al_2O_3 (a) before and (b) after SO_2 treatment and $Pt/Ba/Al_2O_3$ (c) before, (d) after SO_2 treatment, and (e) after H_2 regeneration.

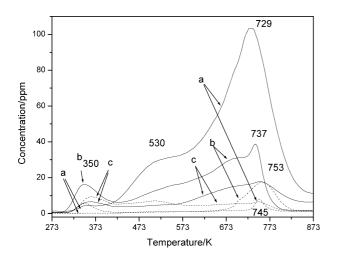


Fig. 2. TPRS of NO (dashed line) and NO₂ (solid line) released from Pt/Ba/Al₂O₃ on sample (a) without SO₂ treatment, (b) after SO₂ treatment in N₂, and (c) after SO₂ treatment in air and then exposure to 750 ppm NO₂ in N₂/air at 673 K and cooled to 298 K, and then heated to 873 K at 4 K min⁻¹ in a 7.98% air/balance N₂ gas mixture (50 cm³ min⁻¹).

NO was released in a single peak at 745 K. For catalysts which had been exposed to SO₂ either in N₂ or in air, the TPRS profiles (Fig. 2b and 2c) were similar to the SO₂-free sample, except that intensities of peaks were diminished for catalysts which had been exposed to SO₂. The exception was the maxima at 350 K which was increased for samples exposed to SO₂ in either air or N₂. Another characteristic of the NO₂ desorption profile was that the peak of maximum intensity was shifted from 729 to 737 K for sample treated in SO₂/N₂ and to 753 K for sample treated in SO₂/air. In the case of the NO profiles, the most significant difference between SO₂-free and SO₂-treated samples were the appearance of a maximum at ca. 350 K for the latter.

Although varying degrees of efficiency were achieved in terms of catalyst regeneration, the shapes of desorption profiles were similarly independent of the reductant used (Fig. 3). NO₂ was the predominant NO_x gas irrespective of the reductant employed. NO was released as the minor component in all cases, at a temperature which corresponded with the maximum in the NO₂ profile. Sample exposed to H₂ (Fig. 3a) gave the most intense NO₂ peak, indicating that regeneration of the NO_x storage capacity had been most successful using this gas. Sample exposed to C₃H₆ gave the least intense NO_2 peak (Fig. 3c), indicating that it was the least effective reductant for NO_x storage capacity regeneration. The NO released was much less than NO2, but also followed the trend that the amount released was greatest following H₂ treatment and lowest after C₃H₆ treatment (Table 1). The peak maxima of the NO₂ profile was shifted from 750 to 725 K for the sample treated with H₂ and CO and to 695 K for the sample treated with C_3H_6 .

 NO_x released (% of total NO_x capacity) from different pretreated Pt/Ba/Al₂O₃ is summarised in Fig. 4. Storage capacity was significantly decreased after exposure to SO₂ in

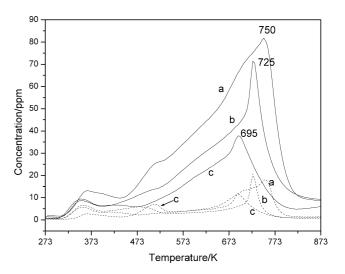


Fig. 3. TPRS of NO (dashed line) and NO₂ (solid line) released from Pt/Ba/Al₂O₃ after exposure to 750 ppm NO₂ in N₂/air at 673 K and cooled to 298 K, and then heated to 873 K at 4 K min⁻¹ in a 7.98% air/balance N₂ gas mixture (50 cm³ min⁻¹) on samples which had been exposed to SO₂ in air at 673 K and then regenerated with (a) H₂, (b) CO, or (c) C₃H₆.

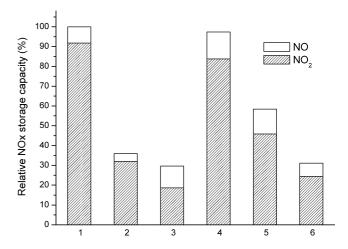


Fig. 4. Summary of the relative amounts of NO_x released from different pretreated $Pt/Ba/Al_2O_3$ samples as a percentage of total NO_x released from fresh sample. (1) Fresh sample, (2) exposed to SO_2/N_2 , exposed to SO_2/air with (3) no regeneration, or (4) regenerated with H_2 or (5) regenerated with CO or (6) regenerated with C_3H_6 .

air (30%) or in N₂ (36%). The storage capacity could be almost completely recovered (97.4%) using H₂ as reductant at 673 K. When this sample underwent repeated cycles of SO₂ exposure/hydrogen regeneration/NO₂ adsorption (up to 5 cycles), the total NO_x uptake never fell below 95% (with respect to the fresh catalyst). The other reductant gases were less efficient with CO regeneration proved to be least effective (31.2%), with recovery being only marginally higher than for catalysts which had not undergone regeneration (30%).

DRIFTS spectra of samples after exposure to NO₂ at 673 K are shown in Fig. 5. Bands at 1432 and 1320 cm⁻¹ are consistent with results for previous studies and were ob-

Table 1					
Summary of NO _x	released	during	TPRS	experimen	ts

Pretreatment	$NO_2 \text{ (mmol/gcat)} (T_{max} \text{ (K)})$				NO (mmol/g _{cat})	$NO_x^a (mmol/g_{cat})$
	Peak 1	Peaks 2, 3	Peaks 3, 4	Total ^a		
Without SO ₂	0.0056 (326)	0.175 (573)	0.257 (735)	(0.438) 0.424	0.038 (760)	(0.475) 0.462
SO ₂ in N ₂	0.0182 (360)	0.081(605)	0.0438 (723)	(0.143) 0.148	0.018 0.0075 (355) 0.0104 (753)	(0.161) 0.166
SO ₂ in air	0.0059 (366)	0.0178 (482)	0.0752 (732)	(0.099) 0.086	0.051 0.0079 (370) 0.026 (500) 0.016 (723)	(0.150) 0.137
With H ₂ regeneration	0.019 (370)	0.156 (580) 0.177 (717)	0.028 (750)	(0.380) 0.387	0.063	(0.443) 0.450
With CO regeneration	0.0145 (352)	0.166 (666)	0.025 (727)	0.210	0.058	0.268
With C_3H_6 regeneration Ba/Al_2O_3	0.0083 (357)	0.0815 (646)	0.0153 (700)	(0.105) 0.113	0.031	(0.136) 0.144
Without SO ₂	0.0306 (356)		0.0748 (760)	(0.1054) 0.107	(0.098) 0.093 0.016 (355) 0.082 (800)	(0.203) 0.200
SO ₂ in N ₂	0.0022 (328)	0.04 (584)	0.09 (775)	0.132	0.074 (793)	0.202
SO ₂ in air	0.0075 (346)	0.0238 (437) 0.0165 (602)	0.0286 (793)	(0.076) 0.054	0.069	(0.145) 0.123
With H ₂ regeneration	0.011 (342)	0.007 (448) 0.017 (604)	0.040 (800)	0.075	0.01	0.085
With CO regeneration	0.018 (352)	0.012 (441) 0.023 (578)	0.055 (795)	(0.108) 0.102	0.009	(0.117) 0.111
With C ₃ H ₆ regeneration HCl-treated Ba/Al ₂ O ₃	0.0015 (337)	0.062 (541)	0.048 (778)	0.111	0.015	0.126
Without SO ₂	0.004 (339)	0.107 (671)	0.065 (746)	0.176	0.03 (786)	0.206
SO ₂ in N ₂				0.111	0.033	0.144
SO_2 in air	0.011 (351)	0.013 (397) 0.032 (652)	0.007 (740)	(0.063) 0.088	0.008 (761)	(0.071) 0.096
With H ₂ regeneration	0.0046 (382)	0.0308 (615)	0.0184 (749)	(0.054) 0.084	0.007	(0.061) 0.091
With CO regeneration	0.0093 (350)	0.015 (390) 0.037 (665)	0.009 (737)	(0.070) 0.092	0.015	(0.085) 0.107

^a NO_x released calculated as the sum of the integrals of individual peaks or () integral over the whole temperature range.

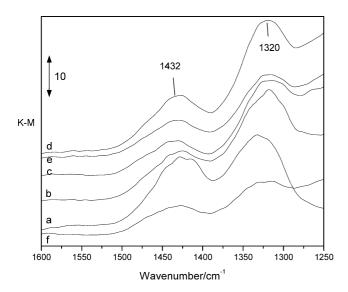


Fig. 5. FTIR spectra of Pt/Ba/Al₂O₃ after NO₂ adsorption at 673 and cooling to 298 K with different pretreated samples. (a) Without SO₂ treatment, (b) after SO₂ treatment at 673 K in N₂, or exposed to SO₂ in air at 673 K with subsequently (c) no regeneration, (d) regeneration in H₂, (e) regeneration in CO, and (f) regenerated in C_3H_6 .

tained whether NO/O_2 mixtures or NO_2 were employed [6, 14,17,19,20]. There were obvious differences in the intensities of this band pair after SO₂ exposure. After exposure to SO_2 in N₂, the intensity of the band at 1320 cm⁻¹ appeared to increase and the component at 1432 cm⁻¹ decreased compared with that of the SO₂-free sample. An increase in relative intensity at 1320 cm⁻¹ was also found for sample exposed to SO₂ in air and again the intensity of the band at 1432 cm^{-1} was decreased. The intensity at 1432 cm^{-1} for the sample after H₂ treatment increased compared to sample without regeneration but without recovering the full intensity of the SO₂-free sample. However, after the same regeneration treatment, the intensity of the band at 1320 cm^{-1} was significantly increased compared to samples either with or without SO₂ exposure (Fig. 5d). Sample treated with CO showed similar trends but to a lesser extent than the effects observed for H₂. For sample regenerated with propene (Fig. 5f), both peaks appeared with slightly lesser intensity than for SO₂-treated (Fig. 5b and 5c) or SO₂-free (Fig. 5a) sample consistent with the lack of storage capacity regeneration using this reductant (Fig. 4).

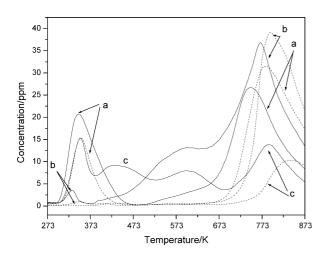


Fig. 6. TPRS of NO (dashed line) and NO₂ (solid line) released from support Ba/Al₂O₃ on sample (a) without SO₂ treatment, (b) after SO₂ treatment in N₂ at 673 K, or (c) after SO₂ treatment in air and then exposure to 750 ppm NO₂ in N₂/air at 673 K and cooled to 298 K and then heated to 873 K at 4 K min⁻¹ in a 7.98% air/balance N₂ gas mixture (50 cm³ min⁻¹).

3.3. Combined TPRS/FTIR experiments: Pt-free Ba/Al₂O₃ catalysts

Fig. 6 shows the NO_x TPRS profiles from Ba/Al_2O_3 which had either been directly exposed to NO₂ (Fig. 6a) or exposed to NO_2 after treatment in SO_2/N_2 (Fig. 6b) or SO₂/air (Fig. 6c). Ba/Al₂O₃ directly exposed to NO₂ (Fig. 6a) released NO₂ at 350, 600, and 745 K while NO was released at 350 and 782 K. For sample after exposure to SO₂ in N2, the ratio of NO/NO2 changed (Table 1) in favour of NO₂. This may have been a consequence of shifting peaks to higher temperature for both NO and NO₂ profiles with virtually no NO_x released at 350 K and a greater NO_2 release at ca. 600 K. Despite the differences in TPRS patterns, the stored NO_x was almost identical to the untreated sample (Table 1). Sample after exposure to SO₂ in air showed a decrease in the overall storage capacity (Table 1), although the peak at 350 K reappeared along with a new feature at 440 K (Fig. 6c). The peak at 600 K was of intermediate intensity between untreated and SO₂ in N₂-treated samples, but the high-temperature peak (788 K) was significantly reduced. NO was released only at the higher of the NO₂ desorption peaks.

NO_x desorption profiles from SO₂-pretreated sample after regeneration with H₂ or CO (Fig. 7a and 7b) were similar in shape to the profile of the TPRS profile with no regeneration (Fig. 6c), with NO₂ maxima at ca. 350, 440, 600, and 793 K. The main difference was the recovery of intensity at ca. 780 K using either H₂ or CO which would account for the partial recovery of the NO_x storage capacity (Table 1). Treatment in propene was less successful in recovering the intensity of the high-temperature feature (780 K) and the partial recovery of overall storage capacity (Table 1) was due to NO_x stored which was released in the mid-temperature range giving maxima at 450, 550, and 650 K (Fig. 7c). A feature observed at ca. 335 K, and also present for H₂ and

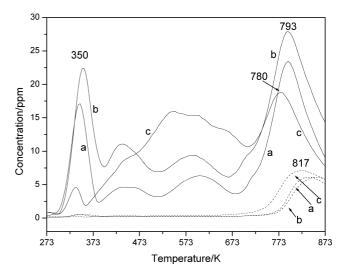


Fig. 7. TPRS of NO (dashed line) and NO₂ (solid line) released from support Ba/Al₂O₃ after exposure to 750 ppm NO₂ in N₂/air at 673 K and cooled to 298 K, and then heated to 873 K at 4 K min⁻¹ in a 7.98% air/balance N₂ gas mixture (50 cm³ min⁻¹) on samples which had been exposed to SO₂ in air at 673 K and then regenerated with (a) H₂, (b) CO, or (c) C₃H₆.

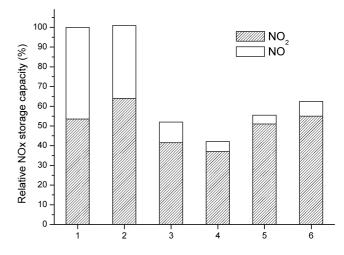


Fig. 8. Summary of the relative amounts of NO_x released from different pretreated supports as a percentage of total NO_x released from fresh sample. (1) Fresh sample, (2) exposed to SO_2/N_2 at 673 K or exposed to SO_2 in air at 673 K and either (3) no regeneration or (4) regenerated with H₂ or (5) regenerated with CO or (6) regenerated with C₃H₆.

CO-treated samples, was observed albeit of lesser intensity. A low intensity peak due to NO was observed at ca. 350 K. The majority of NO was released for all samples in a single peak around 830 K, irrespective of the reductant used to regenerate the catalyst.

 NO_x released as a percentage of the original storage capacity for Ba/Al₂O₃ exposed to different treatments is summarised in Fig. 8. Unlike Pt/Ba/Al₂O₃ exposure of the support to SO₂/N₂ did not reduce the total NO_x storage capacity but did lead to a change in the NO/NO₂ ratio (53.5% NO₂ without SO₂ pretreatment, 64.0% NO₂ after exposure to SO₂ in N₂). Following exposure to SO₂ in air, the storage capacity of Ba/Al₂O₃ was decreased to 52% and subsequent

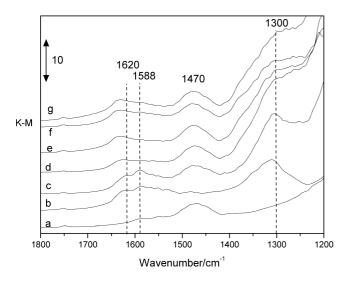


Fig. 9. FTIR spectra of support Ba/Al₂O₃ (a) before NO₂ adsorption, then after exposure to NO₂ at 673 K and cooling to 298 K with different pretreated sample (b) without SO₂, (c) after SO₂ treatment in N₂, or after exposure to SO₂ in air and (d) no regeneration, (e) regeneration in H₂, (f) regeneration in CO, (g) regeneration in C_3H_6 .

 H_2 treatment led to a further decrease (42%). This same effect was not observed for the other reductant molecules with CO treatment reaching 55.5% capacity and propene reaching 62.5% of the original capacity.

DRIFT spectra of Ba/Al₂O₃ following various pretreatments and then exposure to NO2 at 673 K are shown in Fig. 9. The band at 1470 cm^{-1} , also present for the calcined sample (Fig. 9a), is due to BaCO₃. Bands at 1620, 1588, and 1557 cm⁻¹ following exposure to NO₂ (Fig. 9b) are assigned to nitrate ions on the alumina [10,13–15]. The less readily detected 1470 cm⁻¹ band would infer that the formation of nitrates led to at least partial decomposition of BaCO₃, or that the intensity of the broad band at 1310 cm^{-1} , due to nitrate species on alumina [14], obscured the band at 1470 cm⁻¹. The latter feature was clearly observed, however, for a sample which had been exposed to SO₂ in N₂ prior to the NO₂ adsorption (Fig. 9c). There was no obvious change in the intensity of bands assigned to adsorbed NO_x compared to the SO₂-free sample, consistent with the NO_x capacity figures (Fig. 8, Table 1) which showed that SO₂ exposure in N₂ did not influence the storage capacity of the support. Following SO₂ in air (Fig. 9d), the band at 1588 cm^{-1} was significantly attenuated but a shoulder at 1370 cm⁻¹, previously undetected, was apparent. Only the samples which had either no previous contact with SO₂ (Fig. 9b) or were treated in SO₂/N₂ (Fig. 9c) showed a resolved feature at ca. 1310 cm⁻¹. All others showed a continuously rising feature below ca. 1400 cm^{-1} with an indication of the shoulder at 1370 cm⁻¹ (Fig. 9d–9g). The latter has been assigned to $Al_2(SO_4)_3$ [21,22]. In contrast to the fresh sample, a band at 1155 cm^{-1} (not shown) was obvious for the SO₂-exposed samples. Although often used as a characteristic for bulk BaSO₄ [23], the absence of diffraction features [18] for this phase for the Ba/Al₂O₃ (Fig. 1b) suggests

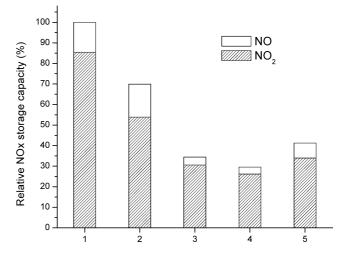


Fig. 10. Summary of the amounts of NO_x released from different pretreated supports as a percentage of total NO_x released from fresh treated support. (1) Fresh sample, (2) exposed to SO₂ in N₂ at 673 K or exposure to SO₂ in air with subsequently (3) no regeneration, or (4) regeneration with H₂ or (5) regeneration with CO.

that this existed as an amorphous phase or at a low concentration. No reductant was capable of recovering the alumina sites responsible for the band at 1588 cm^{-1} (Fig. 9d–9g).

To distinguish effects which arise due to the presence of Pt from the influence of baria dispersion which can be modified by the addition of Pt precursor [14-16], a portion of Ba/Al₂O₃ was treated with HCl solution, washed with ammonia, and calcined.

TPRS profiles for HCl-treated support (not shown) were qualitatively similar to those of Pt/Ba/Al₂O₃. Fig. 10 shows the released NO_x for these samples as a percentage of the storage capacity for the fresh HCl-treated sample. NO_x storage was 0.206 mmol/gcat (Table 1), very similar to the value for the untreated support. Unlike the untreated support (Fig. 8) which was unaffected by a similar treatment, exposure of the HCl-treated Ba/Al₂O₃ to SO₂ in N₂ reduced the NO_x capacity to ca. 70% (Fig. 10). This was qualitatively similar to the result obtained for Pt/Ba/Al₂O₃ (Fig. 4). A clear, enhanced influence on the NO_x storage capacity was observed after exposure to SO_2 in air (Fig. 10), where the NO_x capacity was diminished to only 35% of the original value. As observed for the standard Ba/Al₂O₃ support, HCl-treated support, exposure to SO₂ in air and treatment in H₂ reduced the storage capacity even further (30% of the original). CO, on the other hand, did lead to some level of regeneration (41% of the original).

Absence of peaks at 1750 and 1470 cm⁻¹ in DRIFT spectra of the HCl-treated support (Fig. 11a) confirm that extensive crystalline BaCO₃ was not present possibly due to the well-dispersed barium oxide phase produced after the HCl treatment or that carbonate formation was suppressed due to poisoning of the relevant sites by chloride ions. As for the standard support, bands at 1620, 1588, 1560, and 1310 cm⁻¹ appeared after exposure to NO₂ (Fig. 11b). One difference was a band at 1430 cm⁻¹, due to monodentate nitrate on bar-

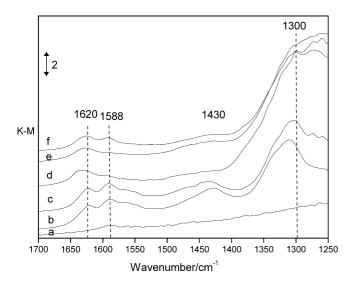


Fig. 11. FTIR spectra of (a) HCl-treated Ba/Al₂O₃ support then after NO₂ exposure between 673 and 298 K for different pretreated samples, (b) without SO₂ treatment, (c) after SO₂ treatment in N₂ at 673 K, and (d)–(f) exposed to SO₂ in air at 673 K with (d) no regeneration or (e) regenerated with H₂ or (f) regenerated with CO.

ium [14,19] which was obscured (Fig. 9a) in the spectrum of the standard support by the band at 1470 cm^{-1} due to the carbonate phase. The band at 1430 cm^{-1} was less dominant while the band at 1300 cm^{-1} was slightly increased following exposure of the sample to SO₂ in N₂ (Fig. 11c). As bands at 1620 and 1588 cm^{-1} due to nitrate on the alumina were largely unaffected by this treatment, the loss in NO_x storage capacity resulting from exposure to SO_2 in N_2 (Fig. 10, column 2) must result from loss of centres on the baria component. The loss of intensity at 1588, 1560, and 1430 cm^{-1} due to prior exposure to SO₂ in air (Fig. 11d) indicates that specific sites on the alumina in addition to sites on baria were poisoned by this treatment which resulted in a greater relative loss in NO_x capacity (Fig. 10). An additional modification to the spectrum involved loss of resolution of the 1300 cm⁻¹ band which was replaced by a rising background below 1400 cm⁻¹ (Fig. 11d). The 1580/1430 cm⁻¹ features were not recovered if SO2 in air was followed by H2 exposure prior to NO₂ adsorption (Fig. 11e); however, the nitrate species on the alumina giving the band at 1588 cm^{-1} could be reestablished to some extent if CO rather than H₂ was used (Fig. 11f).

Fig. 12 shows IR spectra of HCl-treated support exposed to SO₂ in air at 673 K before exposure to NO₂. As above (Fig. 11d), SO₂ treatment in air followed by exposure to NO₂ led to a spectrum where only the band at 1620 cm⁻¹ was apparent, besides the large feature rising from ca. 1400– 1250 cm⁻¹ (Fig. 11d). If the sample showing these features was then heated to 873 K, the 1620 cm⁻¹ band was decreased gradually in intensity and was absent in spectra recorded at 773 K. During heating in air, the feature which had appeared as a rising background below 1400 cm⁻¹ was replaced at around 473 K by a band at 1386 cm⁻¹ (Fig. 12c) which gained in intensity as a function of tem-

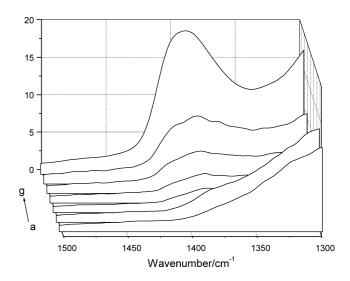


Fig. 12. FTIR spectra of HCl-treated Ba/Al_2O_3 support after SO_2 treatment in air and exposure to 750 ppm NO_2 in N_2/air at 673 K and cooling to 298 K. Sample was heated in 7.98% air/ N_2 mixture to 873 K with spectra recorded at (a) 298, (b) 373, (c) 473, (d) 573, (e) 673, (f) 773, and (g) 873 K.

perature (Fig. 12c–12g), reaching a maximum at 873 K. In the absence of other adsorptive species this suggests transformation of a previously adsorbed form of SO₂.

4. Discussion

Under the conditions employed in this study, the degree of poisoning following SO_2 exposure and the extent of recovery of NO_x storage capacity depend on:

- (i) Whether SO₂ is introduced in air or nitrogen;
- (ii) The presence or absence of Pt;
- (iii) The dispersion/chloride washing of baria;
- (iv) The choice of reductant.

It will become clear in the following discussion that there is considerable interplay between these factors.

4.1. Adsorption and nature of stored SO_x

Consistent with a previous study [24], gaseous O_2 was not essential for SO₂ storage for the Pt-containing catalyst, with similar suppression of NO_x capacity whether SO₂ was admitted in air or in N₂ (Fig. 4). SO₂ uptake is less for catalysts in the absence of Pt [24], although as the overall NO_x storage capacity was also less in the absence of Pt (Table 1), the influence of this smaller amount of adsorbed SO₂ may have a significant effect on the relative amount of residual NO_x capacity. The presence of O₂ was more significant for Ba/Al₂O₃ in the absence of Pt, with NO_x capacity being reduced to ca. one-half of the level obtained for SO₂ in air compared with SO₂ in N₂ (Figs. 8 and 10). Baria dispersion/Cl retention also played a role as SO₂ in nitrogen had no effect on the, as-prepared, Ba/Al₂O₃ sample whereas reduced capacity was found for the HCl-washed support (Fig. 10).

Although not essential [24] for storage over Pt/Ba/Al₂O₃ (Fig. 4), SO₂ uptake is increased by oxygen due to enhanced SO₃ formation [25] which facilitates bulk sulfate formation [24–27]. XRD indicates bulk BaSO₄ formation following exposure of the Pt/Ba/Al₂O₃ catalyst to SO₂ in air (Fig. 1d) but this was not detected for Pt-free sample (Fig. 1b). This is consistent with the diminished SO₂ uptake for Pt-free support [20] and would suggest that in the absence of appropriate conditions necessary to convert SO₂ to SO₃ such as the absence of gaseous oxygen or oxidation catalyst, sulfate formation is limited to the surface layers. NO_x capacity may be equally affected irrespective of whether bulk or surface sulfates are formed since NO_x storage is restricted to the surface baria layers [10,11]. Thus SO₂ in N_2 or in air have similar effects on NO_x storage capacity (Fig. 4) for the Pt-containing catalyst even though bulk sulfate is only formed for the latter treatment (Fig. 1d).

4.2. Influence of adsorbed SO_x on NO_x storage capacity

Similar degrees of efficiency of SO2 in N2 or in air in decreasing NO_x storage capacity could be interpreted as elimination of NO_x storage sites on baria and retention of sites on the alumina since under similar conditions, the latter exhibits ca. 50% of the NO_x storage capacity of a Pt/Ba/Al₂O₃ catalyst [14]. However this is at odds with the FTIR results for this sample (Fig. 5) which show that the dominant bands due to stored NO_x appear at 1432 and 1320 cm⁻¹ due to nitrates on baria [14,17,20]. Nitrates on alumina appearing as a triplet of bands between 1620 and 1560 cm⁻¹ [14,17,20] were not readily detected for Pt/Ba/Al₂O₃ (Fig. 5). However, this was also the case whether the catalyst had been exposed to SO₂ (Fig. 5b and 5c) or not (Fig. 5a), indicating that storage by baria rather than by alumina was the dominant process and that SO₂ exposure reduced the capacity to store NO_x (Fig. 4) by interacting with the former either to poison the surface layer (SO₂ in N₂) or to poison surface sites and produce a bulk sulfate species (SO₂ in air).

The greater relative intensity of bands due to nitrate on alumina (1620, 1588, 1310 cm^{-1}) compared with nitrates on baria (1430 cm⁻¹) for both Ba/Al₂O₃ samples (Figs. 9b and 11b) indicates that the alumina made a proportionally greater contribution to the storage capacity of Ba/Al₂O₃ than for Pt/Ba/Al₂O₃. SO₂ in air led to diminished NO_x capacities for both untreated (Fig. 8) and HCl-washed Ba/Al₂O₃ (Fig. 10) and this treatment led to partial loss of nitrates on the alumina (1588 and 1560 cm^{-1}) and the overall increased intensity below 1400 cm⁻¹. As chelating bidentate nitrate species on alumina (1588 cm^{-1}) are more thermally stable than other forms of nitrate [14,20], including bridging bidentate nitrate (1620 cm^{-1}) which is the least thermally stable [14] and yet still detected after pretreatment with SO₂ (Figs. 9d and 11d), it is clear that these specific adsorption sites on the alumina are poisoned by the SO₂. Burch et

al. [28] noted that SO₂-pretreated Al₂O₃ adsorbed less NO or NO₂, indicating site competition between these species and that a change in desorption temperature occurred for the more strongly adsorbed NO_x. Exposure of the Ba/Al₂O₃ support to SO₂/N₂ led to the appearance of an additional NO₂ desorption feature at ca. 600 K (Fig. 6b) which was not apparent as a resolved feature for the clean support but which was further diminished when SO₂ was admitted in air, with the appearance of an additional component at 440 K (Fig. 6c). It is clear that SO₂ exposure changed qualitative aspects of the stored NO_x as well as the quantitative changes to the storage capacity.

 SO_x species responsible for loss of NO_x capacity on alumina have been described [28,29] in terms of surface sulfates giving bands at 1135 and 1070 cm^{-1} . The latter have been ascribed to a surface sulfite on alumina [30]. These studies [30] also indicated that molecularly adsorbed SO₂, detected as a band between 1330 and 1336 cm⁻¹, was observed following low-temperature (300 K) adsorption. As this molecularly adsorbed form of SO₂ is readily removed at 373 K [30], it is unlikely that it would offer such resistance to NO₂ adsorption at 673 K. Stronger bound forms of SO_x on alumina are produced by reacting SO_2 with oxygen at high temperatures with a band at 1350 [4] or ca. 1380 cm⁻¹ [22,30] being assigned to the ν (S=O) vibration of $Al_2(SO_4)_3$ [4]. The high thermal stability of this species (Fig. 12g) is consistent with behaviour of aluminum sulfate and sulfated alumina decomposition, the latter showing stability up to and above 873 K [2,24], thus supporting an assignment of the 1386 $\rm cm^{-1}$ band to sulfate on the alumina [22]. Aluminum sulfates are less readily formed than barium sulfate [24], consistent with results here where heating Ba/Al₂O₃ from 298 to 873 K (Fig. 12) following exposure to SO₂ in air at 673 K only led to significant evolution of a band at 1386 cm⁻¹ above 673 K reaching maximum intensity at 873 K. Bulk sulfate phases do not show a band at ca. 1380 cm^{-1} [31] and it should be assigned to a surface rather than a bulk sulfate on alumina phase. This is supported by the lack of correlation between this feature and a band at 1035 cm⁻¹ with the latter due to the ν (S–O) mode of the bulk phase sulfate [22]. This assignment is consistent with data for other bulk sulfate phases [31] and in distinguishing surface and bulk sulfates formed on ceria [32]. For engine bench-aged monolith catalysts, it is believed [2] that aluminum sulfate exists as a surface layer. The low relative abundance of the surface aluminum sulfate until reaching elevated temperatures (Fig. 12) would suggest that this species was converted from another mode of stored SO_x , which is more likely to have inhibited NO_x storage as nitrate on the alumina (1588/1560 cm $^{-1}$ bands). As barium sulfate has a greater thermal stability than aluminum sulfate [20], this is not thought to be the source of SO₂ for high-temperature formation of sulfate on alumina (Fig. 12). It is more likely that the high-temperature excursion in air converts surface sulfite species (bands below 1200 cm^{-1}) to surface sulfates (1386 cm^{-1}) and that the former are responsible for lowtemperature site blocking and reduced NO_x capacity of the alumina. In the case of ceria [32], sulfates are produced at the expense of sulfite (or hydrogen sulfite) species leading to the loss of bands below 950 cm⁻¹.

Oxidation of SO₂ to SO₃ is facilitated by Pt and sulfate formation is expected at a lower temperature. Loss of NO_x capacity for Pt-containing sample originated in the main from a loss of baria centres, probably reflecting the ease of barium sulfate formation relative to aluminum sulfate [24]. For reducible oxides such as ceria, sulfate formation may occur from SO_2 in the absence of oxygen [32]; however, it is difficult here to explain why NO_x capacity is reduced when SO₂ in N₂ is admitted to Pt/Ba/Al₂O₃ (Fig. 4) or HCl-treated Ba/Al₂O₃ (Fig. 10). Again it is possible to invoke a role of a lesser oxidised form of SO_x such as sulfite [32] as dominating during SO₂ exposure, and acting to block NO_x storage sites or being converted to a surface sulfate during subsequent exposure of the samples to NO_2 in air, which then competes for adsorption sites with the NO₂. It should be emphasised that under the conditions employed, formation of bulk aluminum sulfate is negligible (Fig. 1), formation of surface aluminum sulfate is limited (below 673 K; Fig, 12), and so loss/recovery of NOx storage capacity does not revolve around these species to any significant extent.

4.3. Influence of reductants on the recovery of NO_x storage capacity

Pt plays a significant role in determining the extent to which NO_x capacity is recovered following SO_2/air and then a brief rich cycle. In the absence of Pt, none of the reductants was effective in recovering storage capacity, irrespective of whether standard or HCl-treated support was employed (Figs. 8 and 10). In fact, post hydrogen exposure in the absence of Pt actually enhanced the poisoning effect of SO_2/air whereas in the presence of Pt, complete NO_x storage capacity recovery was achieved (Fig. 4). The efficiency of the regeneration process for Pt-containing catalyst poisoned at 673 K, $H_2 > CO > C_3H_6$ was the same as that observed for catalysts poisoned at 873 K [2]. In the latter example, samples were exposed to the reductant gas in the absence of oxidant, unlike experiments performed here, and H₂S was evolved [2]. XRD shows that H₂ removed the BaSO₄ phase (Fig. 1e) while FTIR (Fig. 5d) shows that NO_x storage recovery by the baria component occurred. The mechanism of BaSO₄ reduction by H₂ either involves Pt-promoted (hydrogen spillover) reduction to produce BaO, SO₂, and water or BaO, H₂S, and water [2] or sulfate migration from baria to aluminum and hence to the metal site where it undergoes reduction [13]. Additionally, direct sulfate migration from baria to Pt might occur although this would probably be limited to Pt sites at the interface with baria crystallites or in baria-rich regions of the catalyst. Irrespective of the mechanism involved, sulfate reduction by hydrogen may lead to sulphur deposits generated on the Pt under oxygen-deficient conditions [13,27,33]. Note that most regeneration studies of SO₂-treated NSR catalysts involve a rich operating period which involves complete removal of oxygen [2,3,27]. Oxygen in the rich mixture is important as it provides a means of oxidising S deposits on Pt to SO₂ [6,27,33] and under conditions here would appear to prevent formation of a barium sulfide phase [3]. At 673 K oxygen is expected to be fully consumed by H₂ to produce water over these catalysts [20], which is able to hydrolyse S deposits on Pt [9]. As NO storage is less in the absence of Pt (Table 1) [14], full recovery of storage capacity following exposure to SO₂ in air followed by regeneration in H₂/air/N₂ infers that the Pt surface remained S free. Given the greater regeneration efficiency of H₂ compared to CO for Pt/Ba/Al₂O₃ (Fig. 4), we concur [13] that water is a key component in the regeneration of the catalyst.

Water formation from H₂ oxidation is less likely in the absence of Pt and regeneration is expected [13] to be less effective. Both Pt-free supports showed worse NO_x capacity following SO₂/air then H₂/air/N₂ than if the latter treatment were omitted (Figs. 8 and 10). Although barium sulfate may be converted to sulfide in H₂ at temperatures above 623 K giving a peak at $2\theta = 24.1^{\circ}$ in XRD patterns [3], the lack of XRD evidence for a crystalline sulfate phase for the support (Fig. 1b) would make it unlikely that BaS would be detected here. This sulfide phase was stable up to 1073 K in H_2/N_2 [3] but might be expected to convert to a more oxidised state when heating in an oxidising atmosphere. This would be consistent with FTIR data (Fig. 12) where the surface aluminum sulfate phase (1386 cm^{-1}) was detected for the support heated in air above 673 K. An alternative to oxidation of the sulfide phase involved displacement by carbonate in the presence of CO₂ [3]. Ba/Al₂O₃ in the absence of Pt showed activity for CO oxidation at 673 K [34], suggesting that CO₂ may have been a component of the gas phase under these conditions. Note that both support samples showed greater recovery of NO_x capacity following heating in CO compared with H₂ (Figs. 8 and 10). However, FTIR spectra do not show (Figs. 9f and 11f) enhanced carbonate formation (1470 cm⁻¹) following exposure to CO/air compared to hydrogen/air-treated samples (Figs. 9e and 11e) or to the support where no attempt was made to regenerate (Figs. 9d and 11d) and thus displacement of a reduced form of stored S by preferential carbonate formation [3] would not appear to explain the limited (Table 1), but reproducible, improvement in NO_x capacity following treatment in this gas mixture. These same spectra show recovered intensity at 1588 cm⁻¹ following exposure to CO/air (Figs. 9f and 11f) which was not achieved in H2/air (Figs. 9e and 11e) and differences in NO_x capacity following CO/air compared to the H₂/air may originate in the recovery of chelating bidentate nitrate sites [14] on the alumina.

As for higher sulfation temperatures elsewhere [2,3,13], CO was less effective than H₂ in recovering NO_x capacity for the SO₂-treated Pt-containing sample (Table 1, Fig. 4). One could ascribe this difference to the relative ability of the two reductants to interact directly with the sulfate group on baria, which, for the Pt-containing sample, was the principle location of NO_x storage sites (Fig. 5). However, Limousy et al. [13] have shown that unless water is included in the feed gas, CO does not produce H_2S or COS over a sulfated NO_x trap, while in the presence of water, high emissions of H₂S are detected. The latter assumes that in the absence of water, metal sulfides are formed under rich conditions [13,27,33] which prevent subsequent sulfate reduction and recovery of NO_x capacity. A weak band at 2115 cm⁻¹ due to CO adsorbed at oxidised Pt centres was observed following SO2 in air at 673 K and then CO followed by exposure to NO₂. This may reflect the relative strengths of CO adsorbed at oxidised sites compared to Pt⁰ [35] with CO being readily displaced at 673 K by NO₂ in the gas stream rather than indicating blockage of the latter sites by S residues. Amberntsson et al. [27] also failed to observe CO adsorbed on Pt⁰ sites after exposing the catalyst to SO₂ at 673 K, suggesting that poisoning of the Pt sites by S may be significant.

Propene in He at 623 K had negligible effects on IR band intensities of sulfate species [4] and this was the least effective of the three reductant gases used here for the $Pt/Ba/Al_2O_3$ sample. SO_2 on Pt is highly reactive with C_3H_6 resulting in S deposits on Pt [36]. In the absence of SO₂ and under excess oxygen conditions, the catalyst used here is highly active for propene oxidation at 673 K [20]. As outlined above, both water [13] and CO₂ [3] are beneficial in regenerating NO_x storage capacity of SO_2 -poisoned catalysts and it might be assumed that combustion products were not present in significant concentrations under the rich propene/air/nitrogen pulse. Similarly, propene decomposition to yield H₂ was also apparently insignificant given the efficiency of the latter to regenerate SO₂-treated Pt/Ba/Al₂O₃ (Fig. 4). The minor recovery of NO_x capacity for the Pt-free sample following treatment in propene can be attributed, as for CO, to the recovery of sites where NO_x was adsorbed as a chelating nitrate (1588 cm^{-1}) on the alumina (Fig. 9g). A conclusion which could be drawn from the limited success in recovering NO_x storage capacity for the Pt-free samples (Figs. 8 and 10) is that the reductant molecules are inefficient at directly reducing the sulfate or other forms of adsorbed SO_x . This provides some support for a mechanism involving migration of a sulfate unit (or other form of stored SO_x) from baria to the noble metal via the alumina surface [13,25] followed by reduction on the metal site, the retention of an S atom, and eventually deactivation in the absence of water. The alternative that reductant is activated on the metal and transferred to the support might lead to an equivalent scenario if the H₂S released [13] then interacts with the Pt releasing H₂ and retaining S. Experiments to distinguish between the two pathways are in progress [34].

5. Conclusions

Exposure of SO₂ at 673 K leads to loss of NO_x storage capacity due to formation of barium sulfate and other S-containing species on the support. In the presence of Pt, the former makes the greatest contribution to loss of NO_x capacity whereas for the support alone, SO_x adsorption on the alumina makes a significant contribution to lost NO_x uptake. Sulfates on the alumina are only formed in large quantities at higher (T > 673 K) temperatures. In the presence of Pt the order of regeneration efficiency was $H_2 > CO >$ propene while in contrast hydrogen was least effective in the absence of Pt. The presence of oxygen during the rich regeneration period appears be significant as this allows formation of water which has the potential to hydrolyse S residues on the noble metal. The use of hydrogen at low temperatures to release and reduce stored NO_x has the potential to limit sulfate buildup and thus limit the extent to which deactivation due to growth of barium sulfate particles take place and to avoid thermal degradation of the catalyst which accompanies hightemperature desulfation.

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